



**SERVIÇO PÚBLICO FEDERAL
UNIVERSIDADE FEDERAL DE UBERLÂNDIA
INSTITUTO DE BIOTECNOLOGIA
PROGRAMA DE PÓS-GRADUAÇÃO EM GENÉTICA E BIOQUÍMICA**

**BIOMONITORAMENTO DOS EFEITOS GENOTÓXICOS RELACIONADOS À
POLUIÇÃO ATMOSFÉRICA EM AMBIENTES DE INTENSO TRÁFEGO DE
VEÍCULOS: CONTRIBUIÇÕES PARA A VIGILÂNCIA EM SAÚDE AMBIENTAL DE
POPULAÇÕES EXPOSTAS**

Aluno: Carlos Fernando Campos

Orientador: Prof. Dr. Boscolli Barbosa Pereira

**UBERLÂNDIA - MG
2023**



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**Tese apresentada à Universidade
Federal de Uberlândia como parte
dos requisitos para obtenção do
Título de Doutor em Genética e
Bioquímica (Área Genética)**

**UBERLÂNDIA - MG
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PROGRAMA DE PÓS-GRADUAÇÃO EM GENÉTICA E BIOQUÍMICA

ATA DE DEFESA PÓS-GRADUAÇÃO

Programa de Pós-Graduação em:	Genética e Bioquímica				
Defesa de:	Doutorado Acadêmico/ PPGGB				
Data:	Trinta de maio de dois mil e vinte e três	Hora de início:	13:30h	Hora de encerramento:	17:00h
Matrícula do Discente:	11823GBI003				
Nome do Discente:	Carlos Fernando Campos				
Título do Trabalho:	Biomonitoramento dos efeitos genotóxicos relacionados à poluição atmosférica em ambientes de intenso tráfego de veículos: contribuições para a Vigilância em Saúde Ambiental de populações expostas.				
Área de concentração:	Genética				
Linha de pesquisa:	Genética, Epigenética, Biologia e Melhoramento de Plantas e Animais.				
Projeto de Pesquisa de vinculação:	(Bio) indicadores, marcadores e monitores selecionados para estudos em Ecotoxicologia e Saúde Ambiental.				

Aos trinta dias do mês de maio de dois mil e vinte e três, às 13:30 horas, reuniu-se via web conferência pela Plataforma Google Meet, em conformidade com a Portaria nº 36, de 19 de março de 2020 da Coordenação de Aperfeiçoamento de Pessoal de Nível Superior – CAPES, Resolução de nº 06/2020 e Resolução nº 19/2022 do Conselho de Pesquisa e Pós-graduação pela Universidade Federal de Uberlândia, a Banca Examinadora, designada pelo Colegiado do Programa de Pós-graduação em Genética e Bioquímica, assim composta: Dr^a. Juliane Silberschmidt Freitas, Dr. Dieferson da Costa Estrela, Dr. Luis Paulo Pires, Dr. João Vitor Meza Bravo e Dr. Boscolli Barbosa Pereira, orientador do candidato e demais convidados presentes conforme lista de presença. Iniciando os trabalhos o presidente da mesa, Dr. Boscolli Barbosa Pereira apresentou a Comissão Examinadora e o candidato, agradeceu a presença do público, e concedeu ao discente a palavra para a exposição do seu trabalho. A duração da apresentação do discente e o tempo de arguição e resposta foram conforme as normas do Programa de Pós-graduação em Genética e Bioquímica. A seguir o senhor presidente concedeu a palavra, pela ordem sucessivamente, aos examinadores, que passaram a arguir o candidato. Ultimada a arguição, que se desenvolveu dentro dos termos regimentais, a Banca, em sessão secreta, atribuiu os conceitos finais. Em face do resultado obtido, a Banca Examinadora considerou o candidato:

APROVADO.



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Esta defesa de Tese de Doutorado é parte dos requisitos necessários à obtenção do título de Doutor. O competente diploma será expedido após cumprimento dos demais requisitos, conforme as normas do Programa, a legislação pertinente e a regulamentação interna da UFU. Nada mais havendo a tratar foram encerrados os trabalhos. Foi lavrada a presente ata que após lida e achada conforme foi assinada pela Banca Examinadora.



Documento assinado eletronicamente por **Boscolli Barbosa Pereira, Professor(a) do Magistério Superior**, em 30/05/2023, às 16:59, conforme horário oficial de Brasília, com fundamento no art. 6º, § 1º, do [Decreto nº 8.539, de 8 de outubro de 2015](#).



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Documento assinado eletronicamente por **Luis Paulo Pires, Técnico(a) de Laboratório**, em 30/05/2023, às 16:59, conforme horário oficial de Brasília, com fundamento no art. 6º, § 1º, do [Decreto nº 8.539, de 8 de outubro de 2015](#).



Documento assinado eletronicamente por **Dieferson da Costa Estrela, Usuário Externo**, em 30/05/2023, às 17:00, conforme horário oficial de Brasília, com fundamento no art. 6º, § 1º, do [Decreto nº 8.539, de 8 de outubro de 2015](#).



Documento assinado eletronicamente por **Juliane Silberschmidt Freitas, Usuário Externo**, em 30/05/2023, às 17:00, conforme horário oficial de Brasília, com fundamento no art. 6º, § 1º, do [Decreto nº 8.539, de 8 de outubro de 2015](#).



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Ficha Catalográfica Online do Sistema de Bibliotecas da UFU
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C198 Campos, Carlos Fernando, 1989-
2023 Biomonitoramento dos efeitos genotóxicos relacionados
à poluição atmosférica em ambientes de intenso tráfego
de veículos: contribuições para a vigilância em saúde
ambiental de populações expostas [recurso eletrônico] /
Carlos Fernando Campos. - 2023.

Orientador: Boscolli Barbosa Pereira.
Tese (Doutorado) - Universidade Federal de Uberlândia,
Pós-graduação em Genética e Bioquímica.
Modo de acesso: Internet.
Disponível em: <http://doi.org/10.14393/ufu.te.2023.266>
Inclui bibliografia.
Inclui ilustrações.

1. Genética. I. Pereira, Boscolli Barbosa, 1986-,
(Orient.). II. Universidade Federal de Uberlândia. Pós-
graduação em Genética e Bioquímica. III. Título.

CDU: 575

Bibliotecários responsáveis pela estrutura de acordo com o AACR2:
Gizele Cristine Nunes do Couto - CRB6/2091
Nelson Marcos Ferreira - CRB6/3074



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ALUNO: Carlos Fernando Campos

COMISSÃO EXAMINADORA

Presidente: Prof. Dr. Boscolli Barbosa Pereira (Orientador)

Examinadores:

Prof. Dr. Luis Paulo Pires (UFU)

Prof. Dr. João Vitor Meza Bravo (UFU)

Profa. Dra. Juliane Silberschmidt Freitas (UEMG)

Prof. Dr. Dieferson da Costa Estrela (IFTM)

Data da Defesa: 30/05/2023

As sugestões da Comissão Examinadora e as Normas do PGGB para o formato da Dissertação/Tese foram contempladas

Prof. Dr. Boscolli Barbosa Pereira

Dedico este trabalho a meu amigo, professor e orientador Boscolli Barbosa Pereira, à minha irmã, Fabiana Batista Campos e à minha mãe, Maria da Glória Batista Campos.

AGRADECIMENTOS

Agradeço à Universidade Federal de Uberlândia.

Agradeço às agências de fomento FAPEMIG, CNPq e CAPES.

Agradeço ao amigo e professor Nilson Penha-Silva, pelos inúmeros ensinamentos.

Agradeço ao amigo e professor Edimar Olegário de Campos Júnior, pelos inúmeros ensinamentos e diversas colaborações.

Agradeço à professora Ana Maria Bonetti, grande cientista e exemplo de dedicação.

Agradeço ao professor Robson José de Oliveira Júnior.

Agradeço à Vanessa Santana Vieira Santos, por toda ajuda.

Agradeço aos meus familiares Igor Campos Cunha, Matheus Campos Cunha e Sérgio Roberto de Campos pelo auxílio nas coletas.

Agradeço ao meu grande amigo de longa data, Marcelo Reginaldo Campos.

Agradeço aos colegas de laboratório.

Agradeço à minha esposa Liziane Luiz Rodrigues.

Agradeço a todos que contribuíram de forma direta ou indireta para minha formação e contribuições neste trabalho.

Você não pode esperar construir um mundo melhor sem melhorar os indivíduos. Para esse fim, cada um de nós deve trabalhar para o seu próprio aperfeiçoamento e, ao mesmo tempo, compartilhar uma responsabilidade geral por toda a humanidade.

(Marie Curie).

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APRESENTAÇÃO

O uso de transporte individual, que no Brasil está representado pela crescente demanda por carros e motocicletas, é insustentável. Do ponto de vista econômico, os custos de aquisição, manutenção e abastecimento são cada vez maiores para os consumidores. Em contraste, transporte público e alternativas não motorizadas não recebem a prioridade necessária na agenda das políticas públicas. No que diz respeito ao eixo socioambiental, as emissões veiculares, bem como suas consequências para o ambiente e saúde, afetam principalmente a parcela mais vulnerável das populações urbanas.

Ainda que a poluição atmosférica seja um problema de Saúde Pública globalmente reconhecido, os parâmetros ambientais utilizados para monitorar a qualidade do ar são corriqueiramente incompletos, desatualizados e inadequados para a proteção de populações expostas aos poluentes do ar. Nessa direção, as evidências científicas devem subsidiar as políticas e os processos decisórios, incluindo a revisão de parâmetros ambientais. Esses parâmetros, inclusive, devem considerar as variáveis físicas, químicas e biológicas de forma integrada, a partir de programas de biomonitoramento da qualidade do ambiente.

Na presente tese, o tema biomonitoramento é abordado com ênfase nos efeitos genotóxicos relacionados à poluição atmosférica. O trabalho foi organizado em três capítulos, sendo o primeiro capítulo uma revisão crítica da literatura. Os capítulos II e III foram escritos na língua inglesa, trazem os resultados dos experimentos realizados e foram escritos no formato de artigo científico.

A partir dos pressupostos teórico-práticos apresentados no capítulo inicial, avaliamos, nos capítulos II e III, a sensibilidade e viabilidade do emprego do Teste de Micronúcleo em *Tradescantia pallida* como ferramenta para complementar os parâmetros físico-químicos de qualidade ambiental, considerando diferentes condições de tráfego veicular.

CAPÍTULO I

FUNDAMENTAÇÃO TEÓRICA

1 THEORETICAL FRAMEWORK

1.1 IMPACT OF INDIVIDUAL MOBILITY MODEL ON ECONOMY, HEALTH, AND ENVIRONMENT

Population growth results in an increased need for urban mobility. The mobility of people is largely dependent on road transportation, particularly individual motorized transportation. Data from the Institute of Applied Economic Research (IPEA) released in 2022 indicate that over the past 20 years, there has been a growing use of individual transportation at the expense of public transportation systems in Brazilian cities (CARVALHO, 2022).

During this period, the country's economic policy granted tax exemptions for industrial goods and favored the expansion of consumer credit, thus facilitating the purchase and financing of vehicles (MENDONÇA; SACHSIDA, 2014; LUCINDA; PEREIRA, 2017). There were also policies implemented to reduce taxes on automobiles and fuels (VASCONCELLOS, 2018). Despite the economic recession experienced from 2015 onwards, in previous years, an increase in the purchasing power of Brazilian families was observed, especially among middle and low-income families in the North and Northeast regions (PEREIRA et al., 2021). Additionally, there has been an increase in public transportation fares due to inflation. All these factors, combined with longer travel times, reduced comfort, and slower speeds of public transportation, favor the choice of private transportation (LIAO et al., 2020).

On one hand, the inclusion of middle and low-income families in the consumer market for durable goods, including cars and motorcycles, and the increase in purchasing power, along with the growth of the automotive sector, are positive factors for the macroeconomy. However, the increase in individual motorized transportation and the reduction in the use of public transportation have negative consequences for public health and the environment. One of the most immediate effects of the expansion of the car fleet is the deterioration of urban mobility conditions, with worsened traffic congestion, increased travel time, a higher number of accidents, reduction of green areas, and increased levels of air pollution (PEREIRA et al., 2021).

Air pollution is the result of the alteration of the normal physical, chemical, or biological characteristics of the atmosphere, causing harm to humans, flora, fauna, and materials (MILARÉ, 2021). Enhanced by anthropogenic factors, atmospheric alteration caused by pollutants originates from various sources that can be classified as stationary and mobile. Stationary sources mainly result from human activities, such as activities in refineries and petrochemical industries. In turn, mobile sources include motor vehicles, which have the highest pollution potential in this category, producing diffuse emissions that rapidly spread through the atmosphere (DERÍSIO, 2017).

Individual urban mobility is responsible for high levels of air pollutant emissions, mainly due to the use of vehicles powered by fossil fuels, which emit compounds such as carbon dioxide (CO₂), nitrogen dioxide (NO₂), particulate matter (PM), and other pollutants harmful to human health and the environment (FERNANDO; HOR, 2017).

The Global Air Quality Guidelines published by the World Health Organization (WHO) recommend air quality levels for six pollutants internationally recognized as significant contributors for a wide range of air-related diseases: particulate matter, sulfur dioxide (SO₂), carbon monoxide (CO), nitrogen oxides (NO_x), hydrocarbons (HC), and other sulfur oxides (SO_x) (WHO, 2021). Their function is to assess the extent to which the composition of contaminated air deviates from its ideal purity, providing data on the severity levels of air pollution (WHO, 2021).

In Brazil, the Air Quality Index, monitored by CETESB (2021), follows the parameters recommended by CONAMA Resolution No. 491, dated November 19, 2018, which establishes air quality standards in the country. Particulate matter with an aerodynamic diameter of up to 10 µm (PM₁₀) is considered inhalable particulate matter, which is further subdivided into fine fraction PM₁₀, with a size of up to 2.5 µm (known as PM_{2.5}), and coarse fraction PM₁₀, with particles ranging in size from 2.5 to 10 µm (or PM_{2.5-10}).

Constituted by particles suspended in the air, particulate matter (PM) is one of the main agents causing negative effects on the body (GBD, 2018; VANDERLEI et al., 2009) and the environment (CETESB, 2020). Table 1 depicts the harmful effects caused by the major pollutants released into the atmosphere.

Table 1 - Sources and harmful effects of the main vehicular pollutants on the environment and health.

Pollutant	Main sources	Effects on human health	Effects on the environment
MP _{2,5}	Combustion (motor vehicles, secondary aerosol (formed in the atmosphere) such as sulfate and nitrate, among others.	Pathophysiological changes and effect on the body's defense system; MP _{2,5} reaches the pulmonary alveoli and causes irritation, asthma, bronchitis, and lung cancer.	Damage to vegetation, visibility deterioration and soil and water contamination.
MP ₁₀	Combustion reactions (industry and motor vehicles), resuspended dust, secondary aerosol (formed in the atmosphere).	Accumulation in the upper airways, aggravating respiratory problems, such as asthma.	Damage to vegetation, visibility deterioration and soil and water contamination.
Sulfur dioxide (SO ₂)	Fuel oil burning processes, oil refineries, diesel vehicles, pulp and paper production, fertilizers.	Breathing difficulty, changes in the defense system of the lungs, worsening of respiratory and cardiovascular diseases.	Formation of acid rain, corrosion of materials, and damage to vegetation and crops.
Nitrogen dioxide (NO ₂)	Combustion in motor vehicles, industrial processes, thermal plants that use oil or gas, incinerations.	Causes and/or worsens airway problems.	Formation of acid rain, damage to vegetation and crops.
Carbon monoxide (CO)	Incomplete combustion in motor vehicles.	CO acts on the blood, reducing its oxygenation and causing tissue hypoxia. CO can lead to death after a certain period of exposure and concentration.	Increased heat retention in the atmosphere, greenhouse effect.

Source: Adapted from IPEA (2011) and CETESB (2020).

Exposure to high levels of pollutants can cause respiratory, cardiovascular, and cancer-related diseases, and also increase mortality rates. Previous study indicates that air pollution is one of the main causes of increased risk for cardiovascular diseases, ranking as the tenth leading risk factor for global mortality (GBD, 2018).

According to Felin (2018), living in a city with polluted air increases the risk of a heart attack by 75% compared to cities with clean air. Colombini (2008), in a study conducted in the city of São Paulo, identified a significant number of deaths among individuals aged 65 and older related to exposure to PM₁₀. Furthermore, the findings revealed that for every 100 µg/m³ increase in pollutant concentration, there was a 13% increase in overall mortality. Thus, levels of vehicular-related air pollution directly influence air quality, determining the degree and extent of effects on the environment and human health (SAN MARTIN; SAN MARTIN, 2020).

1.2 COVID-19 PANDEMIC AND SOCIAL DISTANCING MEASURES

On December 31, 2019, China alerted the WHO about several cases of unusual pneumonia in Wuhan, a city with 11 million inhabitants located in the central province of Hubei. On January 7, 2020, the identification of a new virus, SARS-CoV-2, was announced (WHO, 2020a). Within a few weeks, the virus spread to several other Asian countries. On January 20, the first case of coronavirus was reported by the United States, and on January 24, the first cases were reported in Europe (HOPKINS, 2020). In Brazil, the first confirmed case of the disease was reported by the Ministry of Health on February 25, 2020 (RODRIGUEZ-MORALES et al., 2020).

On March 11, 2020, after more than 118,000 cases registered in 114 countries, the WHO characterized COVID-19 as a pandemic (WHO, 2020b). Due to the high risk of virus transmission, it became urgent to implement measures and actions to contain the increasing number of COVID-19 cases. European countries, such as Italy, Spain, and the United Kingdom, diverged in their views on the need for social distancing and when to implement it. The main reason for this divergence was associated with the impact of the pandemic on the economy (FERGUSON et al., 2020; XIMENES et al., 2021).

However, the health crisis worsened, and projections generated by mathematical models led to an increasing consensus that social distancing measures would be necessary as the only alternative capable of containing the spread of the pandemic and reducing its effects, particularly severe cases, deaths, and the overload on the healthcare system (FERGUSON et al., 2020; XIMENES et al., 2021).

Importantly, as the cases continued to spread, most countries began implementing restrictions on trade, transportation, and cultural activities. Schools, universities, workplaces, and commercial or religious establishments were closed (DANTAS et al., 2020). As a result, the coronavirus outbreak led to a drastic reduction in vehicular traffic and, consequently, had a positive impact on air quality (SAADAT; RAWTANI; HUSSAIN, 2020).

1.3 IMPACTS OF REDUCED VEHICULAR TRAFFIC ON AIR QUALITY

In several cities, especially in large urban centers, levels of air pollutants such as nitrogen dioxide (NO₂) and fine particles dropped drastically during periods of social isolation and lockdown, emergency protocols implemented to prevent the circulation of people for non-essential activities. In terms of the impact on greenhouse gas emissions generated by the combustion of fossil fuels, projections estimated a significant reduction due to the contraction of economic activity and the decrease in car traffic, which represent one of the main sources of air pollution in urban areas, mainly during periods of greater restrictions on the movement of people (SAN MARTIN; SAN MARTIN, 2020).

Indeed, in 2020, air monitoring studies conducted by NASA revealed a significant reduction in nitrogen dioxide (NO₂) levels, particularly in China, as a result of social distancing measures implemented to contain the virus spread (NASA, 2020a,b). On a global scale, data from the International Energy Agency (IEA) indicate a decline of one million tons of CO₂ per day during this period, mainly due to reduced coal and oil consumption (IEA, 2020).

According to the World Meteorological Organization, Southeast Asia recorded a 40% decrease in harmful airborne particulate matter caused by traffic and energy

production in 2020 (WMO, 2020). In Brazil, satellite images taken in March and April 2020 showed a significant reduction in particulate matter in the metropolitan regions of São Paulo and Rio de Janeiro. Similar results were found in countries such as China, Italy, Spain, and France (MUHAMMAD et al., 2020; NASA, 2020a,b).

Notably, the decrease in air pollution levels and improvement in air quality had a positive impact on population health. Studies have reported changes in hospitalization patterns for certain conditions, including cancer, cardiovascular diseases, and respiratory disorders, with a reduction in hospitalizations for asthma and chronic obstructive pulmonary disease (COPD) (ARAÚJO-FILHO et al., 2020; FONSECA et al., 2021; NORMANDO et al., 2021).

In Brazil, a national observational study reported a decline in hospitalization rates and in-hospital mortality due to respiratory diseases - except for COVID-19 - in the first eight months of the pandemic in the country as a result of measures to contain the virus spread (ALBUQUERQUE et al., 2023).

Thus, it is crucial to monitor air quality to identify changes in air pollution resulting from the implemented restriction measures to contain the COVID-19 pandemic, especially regarding the effects of anthropogenic pollutants.

1.4 MONITORING THE EFFECTS OF AIR POLLUTION CHANGES IN THE URBAN ENVIRONMENT

The increase in population in urban areas, coupled with industrial activity and fossil fuel combustion caused by vehicles, has led to higher doses of complex mixtures of pollutants in the atmosphere, including mutagenic and carcinogenic compounds. Therefore, the use of environmental quality monitoring techniques is required (WHO, 2021).

Despite significant advances in strategies to reduce air pollutant emissions in recent years, air pollution in urban areas remains a serious environmental and health concern. In this regard, the air quality index and ambient concentrations of major air pollutants can be measured through physical-chemical methods and mathematical models (KLUMPP et al., 2006). The data obtained indicate whether the limit values

established by global organizations or recommended by local environmental laws are being respected (CAMARA, 2020).

However, the results of these assessments do not guarantee conclusions about the impact of air pollutants on living organisms. As a consequence, biomonitoring techniques are used to understand the effects of pollutants on the environment, aiming to identify environmental risks that threaten the balance and health of organisms (KOCH et al., 2016; SILVEIRA et al., 2021).

Biomonitors are organisms that accumulate contaminants in their tissues, thereby providing information about the quantitative aspects of environmental quality (HATJE, 2015). Living organisms, or communities of living organisms, can be classified as monitors because they accumulate one or more elements or compounds from the environment, responding simultaneously to different stressors and showing effects on morphological, histological, cellular structures, metabolic processes, behavior, and/or population structure. Moreover, they can reflect stress or the conservation status of a particular environment (KOCH et al., 2016).

In this sense, organisms such as plants, fish, insects, amphibians, birds, mammals, and invertebrates can be used as bioindicators, and their responses can be measured as biomarkers (ARZUMANYAN et al., 2022; GALLITELLI et al., 2022; MURTHY et al., 2022; PARRA-LUNA et al., 2020).

Certain plant species, such as *Tradescantia* sp., *Vicia faba*, and *Allium cepa*, have been employed in biomonitoring research of atmospheric quality. Besides their rapid development and easy propagation, factors such as sensitivity, high efficiency, and low operational cost make these plants ideal bioindicators for monitoring and laboratory or *in situ* investigations (RODRÍGUEZ et al., 2015).

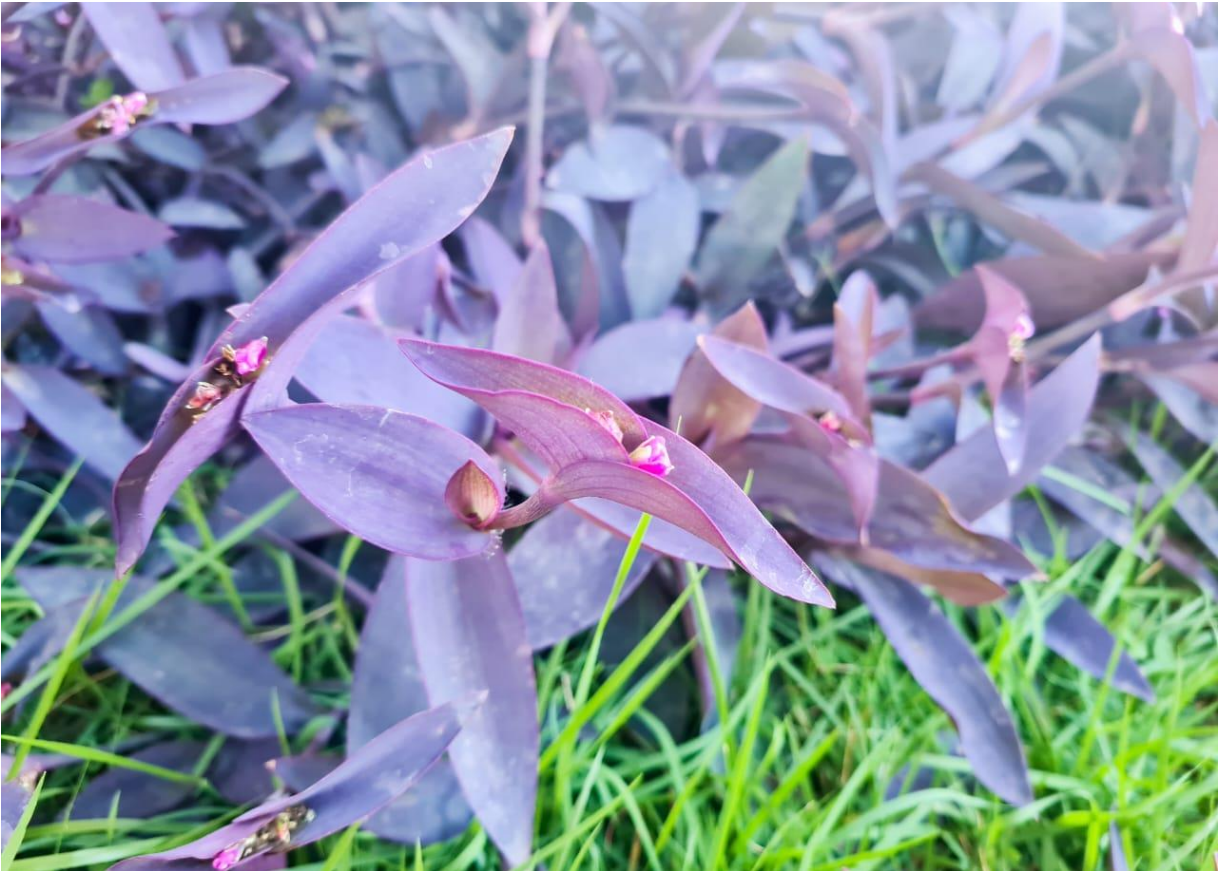
1.5 *Tradescantia* GENUS IN ENVIRONMENTAL BIOMONITORING

Belonging to the family Commelinaceae, the genus *Tradescantia* comprises over 500 species primarily found in tropical and subtropical regions. Some of these species and their clones are used as bioindicators of genotoxicity, for example, *Tradescantia pallida* (FADIC et al., 2016).

Tradescantia pallida (Rose) D.R. Hunt is a small herbaceous plant, measuring 15 to 25 cm in length (Figure 1). Its leaves are fleshy and glabrous, with an epidermis rich in anthocyanin pigments, giving the species a pink or purple coloration. *T. pallida* produces typically solitary flowers throughout the year, in pots, gardens, flowerbeds, and fields, showing easy adaptation and multiplication in any environment (LORENZI, 2015).

T. pallida has six pairs of large and easily observable chromosomes, and cells from almost all parts of the plant - from the root tip to the pollen tube - are continuously developing, favoring its use in cytogenetic and environmental biomonitoring studies (MA; GRANT, 1982). In this regard, the most prominent technique performed with plants of the *Tradescantia* genus is the micronucleus assay (Trad-MCN) (MA et al., 1994).

Figure 1 – *Tradescantia pallida* (Rose) D.R. Hunt



Source: The author (2023).

1.5.1 Trad-MCN as a sensitive biomarker

The micronucleus test in *Tradescantia* was first developed by Ma et al. (1978). This cytogenetic assay is based on a series of procedures for exposing (in the laboratory or *in situ*) *Tradescantia* plants to contaminants, followed by the analysis of micronucleus frequency resulting from chromosomal breakage in meiotic pollen mother cells at the tetrad stage. The tetrad-stage cells are obtained from young inflorescences, which is the ideal period for micronucleus observation, as the cells are in interphase, which in the *Tradescantia* genus lasts for 36 to 48 hours (RODRIGUES; PIMENTEL; WEINSTEIN et al., 1998; GERAS'KIN, S.; EVSEEVA, T.; OUDALOVA, 2011; PEREIRA; CAMPOS JÚNIOR; MORELLI, 2013).

Micronuclei (MN) are structures that result from whole chromosomes or chromosomal fragments that are lost during cell division, remaining in the cytoplasm of interphase cells. Therefore, they result from structural damage and aneuploidy, allowing the detection of clastogenic and aneugenic agents, respectively (RODRÍGUEZ et al., 2015).

The Trad-MCN test is considered a valuable tool by several researchers due to the simplicity of the methodology and the sensitivity of the *Tradescantia* sp. genus to genotoxic agents. In fact, these species have demonstrated precision and efficacy in analyzing the genotoxic potential of air pollutants through the bioassay (MA et al., 1994; GUIMARÃES et al., 2000).

1.5.2 Cytogenetic analyses for micronucleus observation

The steps of selection, fixation, and preservation of young inflorescences used in the micronucleus test with *Tradescantia* are carried out according to the protocol proposed by Ma (1981). In atmospheric biomonitoring studies, this protocol establishes that the inflorescences should be collected and fixed in a solution of acetic acid and ethanol (1:3) for 24 hours (PEREIRA; MORELLI, 2013). Subsequently, the inflorescences should be transferred to 70% ethanol and kept at 6 °C until the cytogenetic analysis.

The preparation of slides is performed by selecting and dissecting the inflorescences, followed by the isolation of floral buds. Initially, intermediate-sized buds from each dissected inflorescence are used to increase the chances of visualizing floral buds with meiotic pollen mother cells at the tetrad stage. The selected floral bud is transferred to a slide and dissected with a histological scalpel to expose the anthers. After the resulting cellular fragments from the dissection of the floral bud are removed and discarded, the anthers are macerated by adding a drop of 2% acetic carmine and covering the slide with a coverslip, gently pressing the anthers to release the cells. Heating for 5 seconds at a temperature of 60 °C is performed to fix the stain in the cells.

Before analyzing the slide, it is recommended that the researcher verify the presence of an adequate number of meiotic pollen mother cells at the tetrad stage. Ma

(1981) establishes the analysis of at least 300 tetrads per inflorescence. The number of micronuclei should be estimated in at least five inflorescences per monitored location or situation. The analyzed slides are observed under an optical microscope with a magnification of 400 times.

The slides should be coded, and the analyses should be performed in a blind study. Structures that measure approximately 1/3 to 1/5 of the main nucleus, exhibit similar chromatin staining and distribution to the nucleus, and are disconnected from it are considered micronuclei.

1.6 HYPOTHESIS AND AIM OF THE STUDY

In this study, we aimed to evaluate the plausibility of employing a biomonitoring program using *T. pallida* in urban places of intense traffic, for periodic assessments of environmental quality. For this purpose, as biomarkers of genotoxicity in response to exposure to an environment of intense vehicular traffic, the frequency of micronuclei and pollen abortivity in inflorescences collected at different intersections with gradual levels of traffic volume were evaluated. The concentrations of bioaccumulated heavy metals in the leaves of the collected plants were also investigated. Furthermore, to test the sensitivity of the proposed biological assessment model, these biological responses were correlated to the environmental variables (i) traffic volume and (ii) concentration of particulate material.

Posteriorly, we focused on the use of *T. pallida* as a biomonitor in situations where traffic varied in the same location, due to business closure measures implemented as a strategy to face the COVID-19 pandemic. We tested the hypothesis that indicators of genotoxicity in *T. pallida* respond significantly to changes in vehicular traffic in the same location. To test our formulated hypothesis, we collected data on vehicular traffic, particulate matter, heavy metals in the soil and plant, besides the genotoxicity analysis using the micronucleus test.

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CAPÍTULO II

ARTIGO CIENTÍFICO EXPERIMENTAL

Título:

Analysis of genotoxic effects on plants exposed to high traffic volume in urban crossing intersections

Periódico:

Chemosphere (Elsevier)

Fator de impacto: 8.943

1 **Analysis of genotoxic effects on plants exposed to high traffic volume in**
2 **urban crossing intersections**

3 Running title: **Genotoxic effects on plants exposed to high traffic**

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24
25 **Abstract:** A biological assessment of environmental quality was performed using the
26 tropical plant species *Tradescantia pallida* (Rose) D.R. Hunt. var. *purpurea* exposed to
27 different levels of air contamination in urban intersections with high volume of vehicle
28 traffic. Air quality (average daily levels of particulate material in the PM_{1, 2.5, 10} fractions)
29 and traffic volume in crossing intersections were monitored for 30 days before the
30 collection of plants. Frequency of micronuclei and pollen abortivity in inflorescences
31 collected at different intersections with gradual levels of traffic volume were evaluated as
32 biomarkers of genotoxicity. In addition, the concentrations of bioaccumulated heavy
33 metals in the leaves of the collected plants were also investigated. The proposed
34 biological assessment model found a positive association between the environmental
35 variables (traffic volume; concentration of particulate material) and biological effects
36 (leaf concentration of Cr and Cd; micronucleus frequencies and pollen abortivity).

37
38 **Keywords:** Urban gardens; Biomonitoring; Genotoxicity; Heavy metals; Micronuclei;
39 Particulate material.

40

41 **Introduction**

42

43 The continuous growth of the motor vehicle fleet has remarkable impacts on the
44 urban environment and health of citizens (Van Veldhoven et al. 2019; Oliveira et al. 2019).
45 In addition to the traffic accidents, congestion and the resulting increase in the daily travel
46 time of people, the increase in the fleet generates emissions of pollutant compounds in
47 the urban atmosphere (Amato-Lourenco et al. 2016; Pereira et al., 2019).

48 Physical-chemical analysis associated with mathematical models allow
49 measurements of environmental concentrations of the major atmospheric pollutants to
50 verify whether limit values established by local environmental laws or recommended by
51 world organizations have been respected (Montoya et al. 2020). Notwithstanding,
52 although air quality is assessed through physical-chemical parameters, which accurately
53 estimate the concentration of pollutants, in fact the obtained results do not guarantee
54 conclusions regarding the impact of these contaminants on living organisms (AL-Alam
55 et al., 2019).

56 In this perspective, the biological assessment of environmental quality has been
57 investigated using plant species that accumulate contaminants (Mahapatra et al. 2019).
58 Bioindicator organisms are important tools in environmental monitoring studies (Ramić
59 et al. 2019). Remarkably, they are able to exhibit alterations in biomarkers, even when
60 exposed to low levels of contamination of the environment (Sinha et al., 2014; Qarri et
61 al. 2019).

62 Thus, the biological assessment of exposure to contaminated air using sensitive
63 plant species to environmental changes is appropriate to detect and early monitor effects
64 that can extend to human health (Placencia et al., 2019; Mišík et al 2019).

65 Several studies carried out with plants of the genus *Tradescantia* have elucidated
66 genotoxic effects of exposure to air pollution in urban environments of intense traffic
67 (Sposito et al. 2017; Rocha et al. 2018). In these assays, researchers generally cultivate
68 clones of *Tradescantia* species that occur naturally in temperate regions and then expose
69 the plants for specific periods to the environment to be investigated. This evaluation
70 model ensures isogenicity, hence avoiding the bias that genetic variability could induce,
71 but the stress caused by displacement and change of environment also interferes on the
72 observed responses (Pereira; Campos Júnior; Morelli, 2013).

73 Alternatively, in the present study, a biological assessment model was employed
74 using the species *Tradescantia pallida*, which has been used successfully in assessing
75 environmental quality in tropical regions (Pereira et al. 2014; 2017; Nakazato et al. 2018).
76 In addition, the evaluations were performed on plants that were already present in public
77 flowerbeds of urban intersections.

78 In this sense, the aim of this research was to evaluate the plausibility of employing
79 a biomonitoring program using plants of the species *T. pallida* in urban places of intense
80 traffic, for periodic assessments of environmental quality. For this purpose, as biomarkers
81 of genotoxicity in response to exposure to an environment of intense vehicular traffic, the
82 frequency of micronuclei and pollen abortivity in inflorescences collected at different
83 intersections with gradual levels of traffic volume were evaluated. Furthermore, the
84 concentrations of bioaccumulated heavy metals in the leaves of the collected plants were
85 also investigated. Additionally, in order to test the sensitivity of the proposed biological
86 assessment model, these biological responses were correlated to the environmental
87 variables (i) traffic volume and (ii) concentration of particulate material.

88

89

90 MATERIAL AND METHODS

91

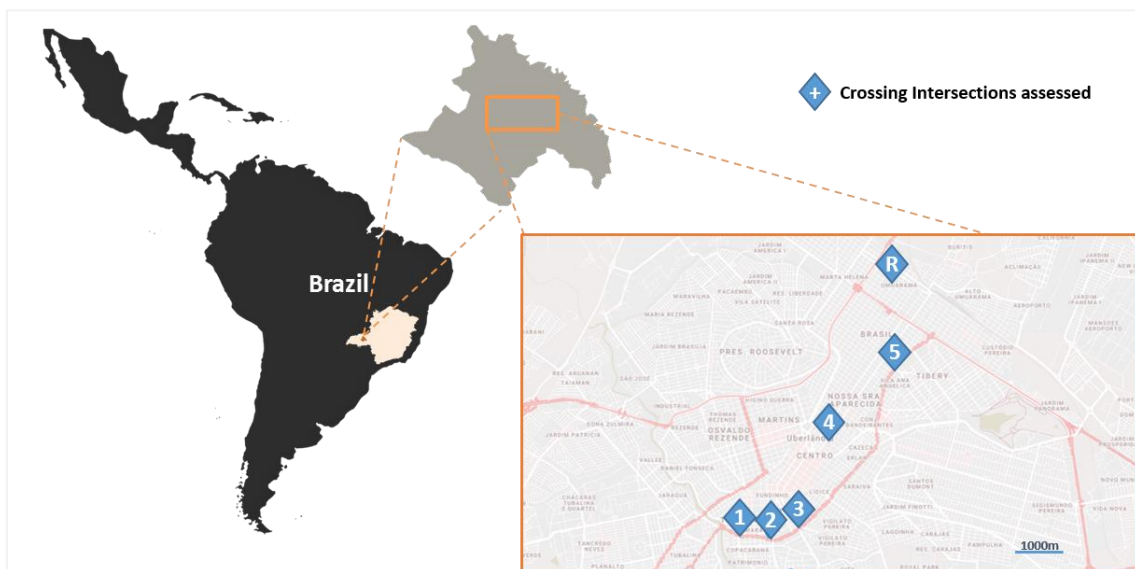
92 Collection sites

93

94 Plants of the species *Tradescantia pallida* (Rose) Hunt. cv. *purpurea* Boom which
95 were present in the flowerbeds of different intersections in the city of Uberlândia, Minas
96 Gerais, were collected. The collection sites were chosen based on crossings with different
97 volumes of daily traffic, but also with an equivalent proportion of light and heavy fleet of
98 motor vehicles. According this criterion, 40 cm stems with young fluorescences of *T.*
99 *pallida* were collected in 5 intersections of intense traffic along the major avenues of the
100 city. In addition, *T. pallida* stems found in the garden of Federal University of Uberlândia
101 were also collected and this site was selected as a reference due to the low volume of
102 vehicular traffic. Figure 1 shows the location of the intersections and the reference site.

103

104 **Figure 1.** Location of the crossing intersections and the reference site assessed.



105

106 **Note:** Intersection 1 (-18.93113, -48.28966); Intersection 2 (-18.93196, -48.28358);
107 Intersection 3 (-18.92986, -48.27775); Intersection 4 (-18.89958, -48.2586); Intersection
108 5 (18.9132, -48.27214) and Reference site (-18.88295, -48.25907).

109

110 **Street-level traffic data and particulate material sampling**

111

112 To assess the exposure conditions of plants to vehicle traffic and the emissions of
113 pollutants, the crossing intersections and the reference site were monitored for 30 days
114 before the collection of plants. The traffic volume was estimated based on video
115 monitoring data provided by the Traffic Department of the city of Uberlândia. The traffic
116 volume (vehicles/day) was estimated by the weighted average of vehicles that passed
117 through the roads (in all directions) considering the proportion of working days and
118 weekends. The average daily levels of particulate material in the PM_{1, 2.5, 10} fractions at
119 each evaluated site was obtained using a portable sampling device (Dusttrak DRX
120 Aerosol Monitor), calibrated for detection of aerosol in sampling volume fixed at
121 3.0L/min.

122

123

124 **Trad-MN assay and Pollen abortivity test**

125

126 To perform the Trad-MN assay, 5 to 10 young inflorescences collected from each
127 collection site were used. According recommendations by Ma et al. (1994), the biological
128 material was fixed in 1: 3 acetic acid to 70% ethanol solution for 24 h and preserved in
129 70% ethanol. Considering the same inflorescences used for the Trad-MN assay, flower
130 buds were fixed in a glacial acetic acid and ethanol (96%) solution (1: 3) for 24 hours and
131 preserved in 75% ethanol. The procedures of excision, staining and analysis were
132 performed according to the protocol of Solenská, Micieta, Misík (2006). For each
133 evaluated site, 10 slides were prepared for counting micronuclei and pollen abortion
134 events. For Trad-MN assay, 300 tetrads per slide were examined; for the Pollen abortivity

135 test, 3000 pollen grains per slide were assessed. All analyzes were conducted using an
136 optical microscope at 400 x magnification.

137 **Bioaccumulation**

138

139 The average leaf concentrations of heavy metals in *T. pallida* cv. *purpurea*
140 collected from different study sites (Kruskal Wallis and Dunn) were obtained according
141 to methodology already described by Campos et al. (2016). The analysis of the heavy
142 metals Lead (Pb), Chromium (Cr), Nickel (Ni) and Cadmium (Cd) in the plant leaves was
143 carried out by using atomic absorption spectrometry in a graphite oven (GFAA),
144 according to the method 7010-USEPA. For the analysis of Cobalt (Co), Barium (Ba),
145 Copper, (Cu), Iron (Fe), Sodium (Na) and Zinc (Zn), the analytical method applied was
146 optical emission spectrometry with argon plasma (ICP-OES), according to protocol
147 established in USEPA method 6010C (USEPA, 2007).

148

149 **Statistical analysis**

150

151 Initially, the Shapiro-Wilk test was applied to assess the normality of all variables
152 of the study. The Particulate Matter (PM) concentrations and the frequencies of
153 micronuclei and abortions in inflorescences of *T. pallida* were compared between the
154 different sites and statistically evaluated using the Kruskal-Wallis and Dunn tests. The
155 differences between the average concentrations of bioaccumulated heavy metals in the
156 leaves of the plants collected at each crossing intersection were subjected to one-way
157 Analysis of Variance (ANOVA) and then compared to reference site using the Tukey test.
158 Additionally, in order to determine correlation patterns between all studied variables that
159 reported a significant difference between the evaluated sites, a multivariate analysis was

160 performed by application of canonical correlation analysis. The statistical model was
161 adjusted to analyze/compare two blocks of variables in an integrated way, as following:
162 environmental (PM_{1, 2.5, 10} fractions; traffic volume) and biological (micronuclei and
163 pollen abortion frequencies; bioaccumulation of metal). P values <0.05 were considered
164 statistically significant for all analyzes.

165

166 **Results and discussion**

167

168 Figure 2 highlights the gradient in the volume of motor-vehicle traffic that passes
169 through the evaluated intersections on a daily basis, revealing the increase of vehicle
170 traffic flow from the peripheral neighborhoods to the center. In addition, it is noteworthy
171 that the levels of particulate material in all assessed fractions (PM_{1; 2.5; 10}) were
172 significantly higher at the intersections in relation to the reference site.

173 Micronucleus frequencies (Figure 3A) and the percentage of pollen abortivity
174 (Figure 3B) in the inflorescences of *T. pallida* collected at the intersections were also
175 significantly higher in comparison to those obtained at the reference site.

176 Also, according to the results of the chemical digestion of the leaves collected at
177 the different crossing intersections, the Cd and Cr concentrations were significantly
178 higher in the samples of the intersections in comparison to the reference site, hence
179 indicating the sensitivity of the species in bioaccumulate these heavy metals. The
180 concentrations of Ni, Pb, Co, Ba, Cu and Zn in leaf samples collected at the intersections
181 and reference site did not differ from each other (Table 1).

182 Table 1. Bioaccumulation of metals in *T. pallida* leaf samples collected at the studied intersections and Reference site.

SITES	[Metals (mg.kg ⁻¹)]							
	Cd	Ni	Cr	Pb	Co	Ba	Cu	Zn
Intersection 1	0.62 ± 0.34b	3.07 ± 2.40a	1.99 ± 1.01b	1.26 ± 0.83a	0.32 ± 0.13a	82.30 ± 62.9a	14.22 ± 8.34a	79.54 ± 19.63a
Intersection 2	0.60 ± 0.24b	3.01 ± 2.22a	2.22 ± 0.80b	1.15 ± 0.65a	0.30 ± 0.12a	77.33 ± 59.8a	13.71 ± 5.85a	91.34 ± 21.33a
Intersection 3	0.55 ± 0.31b	3.03 ± 1.64a	2.01 ± 0.69b	1.23 ± 0.59a	0.28 ± 0.10a	69.71 ± 72.2a	12.21 ± 5.77a	69.83 ± 23.72a
Intersection 4	0.47 ± 0.27b	3.19 ± 2.11a	1.44 ± 0.68b	1.30 ± 0.88a	N.D.	85.08 ± 48.5a	13.84 ± 6.26a	89.12 ± 25.98a
Intersection 5	0.45 ± 0.18b	2.58 ± 2.30a	1.41 ± 0.73b	1.11 ± 0.70a	0.43 ± 0.25a	69.98 ± 66.3a	11.02 ± 4.90a	77.63 ± 23.73a
Reference site	0.12 ± 0.09a	2.04 ± 1.87a	0.35 ± 0.21a	0.81 ± 0.49a	N.D.	71.10 ± 56.4a	10.13 ± 6.11a	80.67 ± 30.41a

183 Asterisk indicates significant difference from Reference site (ANOVA; Tukey; p < 0.05). N.D. Not detected.

184

185 Table 2. Results of Canonical Correlation Analysis (CCA) between environmental and biological variables.

186

Covariates (environmental variables - X_i)	Canonical Variables		Dependent variables (biological variables - Y_i)	Canonical Variables	
	χ_1	χ_2		η_1	η_2
Traffic (X1)	2.2892	-0.3875	Cr (Y1)	1.0896	-0.4242
PM ₁ (X2)	2.2953	7.1199	Cd (Y2)	-1.2198	-1.3695
PM _{2.5} (X3)	-2.7789	-5.8082	MN (Y3)	1.8741	3.7913
PM ₁₀ (X4)	-1.0581	-1.4405	Abort (Y4)	-0.9114	-2.5543
Canonical R	0.9968	0.9220			
p-value	0.0001	0.0318			

187

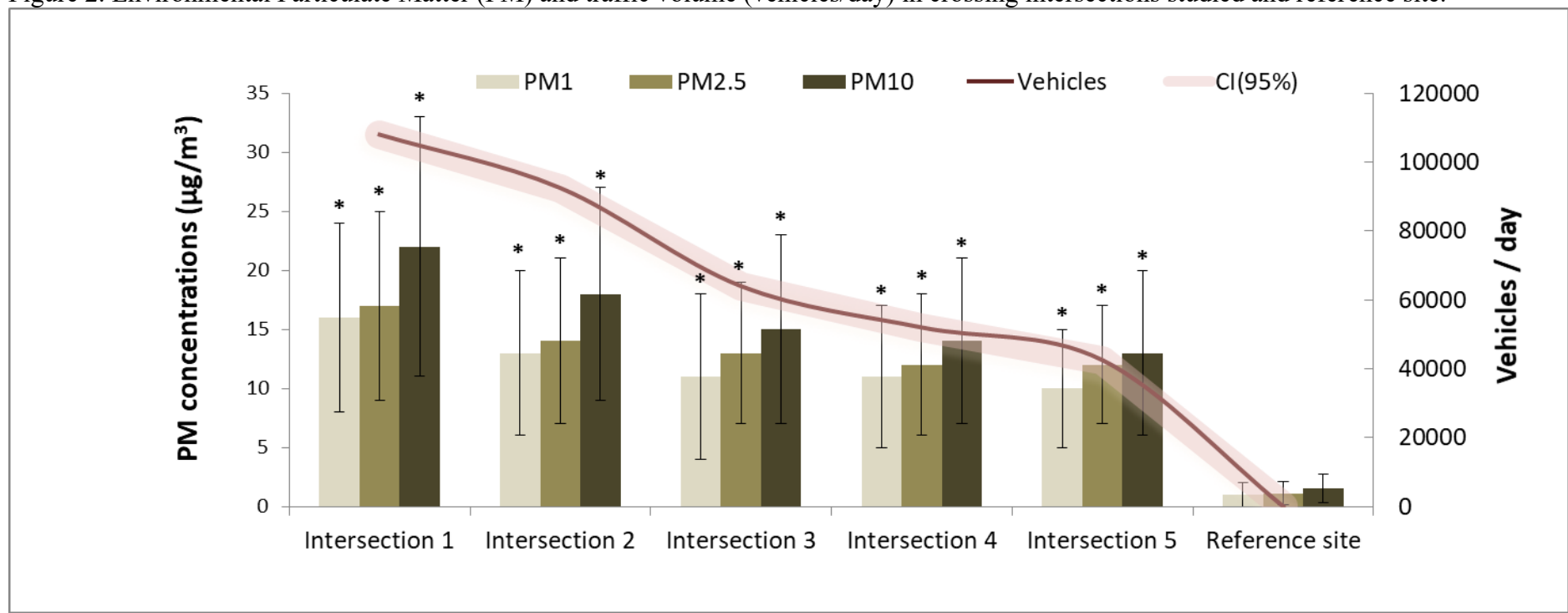
188

189 Table 3. Matrix of correlations (R) between Xi and Yi variables.

	Cr	Cd	MN	Ab
Frota	0.929*	0.910*	0.953*	0.923*
PM1	0.850*	0.860*	0.911*	0.884*
PM2.5	0.871*	0.893*	0.916*	0.920*
PM10	0.747**	0.801*	0.810*	0.782*

190 Significant p value (p<0.05*; p<0.05**), according to Pearson's correlation test.

191 Figure 2. Environmental Particulate Matter (PM) and traffic volume (vehicles/day) in crossing intersections studied and reference site.

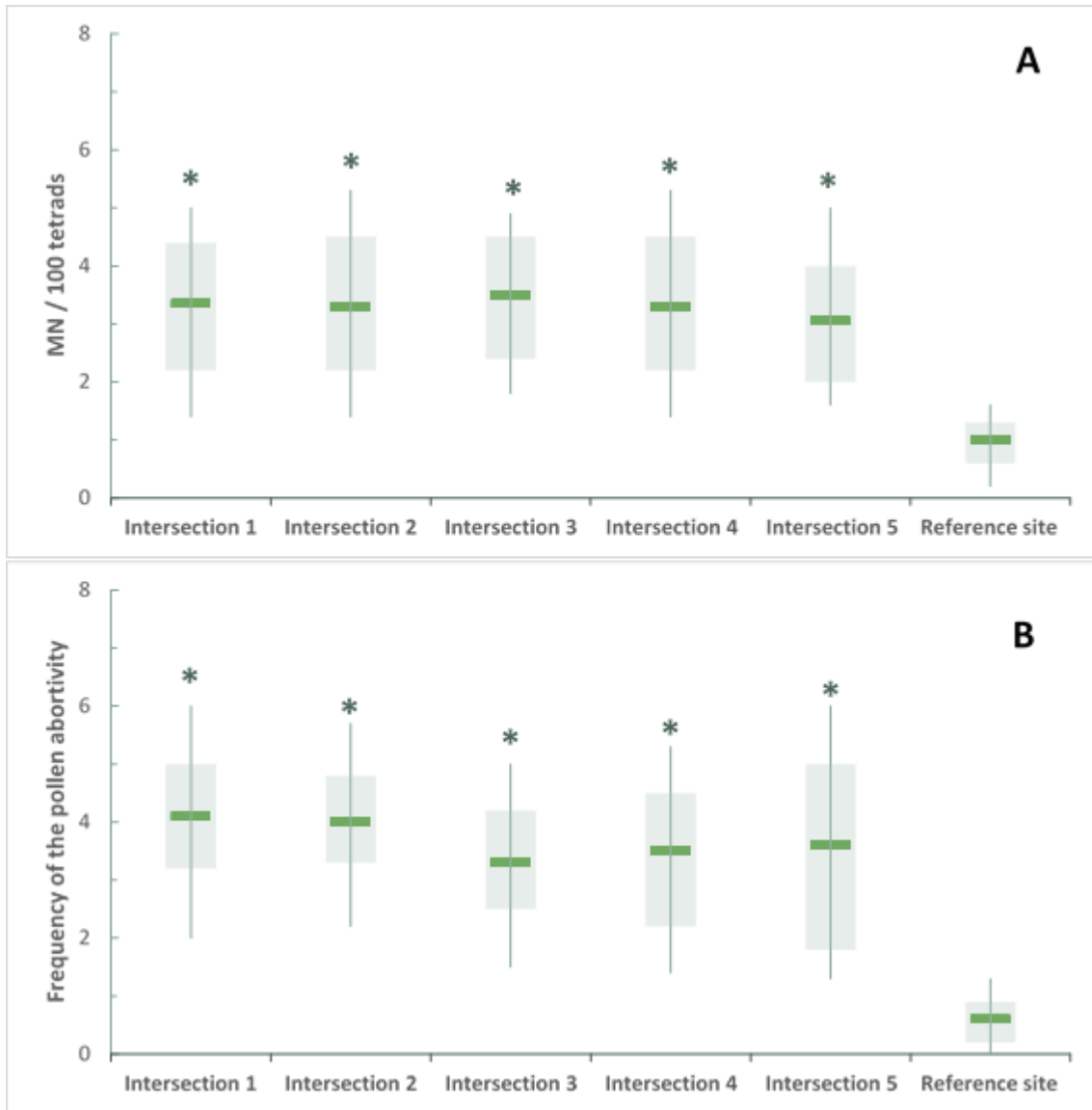


192 Asterisk indicates significant difference from Reference site (Kruskal-Wallis; Dunn; p < 0.05).
 193

194 Figure 3. Frequency of the micronuclei (A) and pollen abortivity (B) in *Tradescantia pallida*
195 collected in crossing intersections studied and reference site.

196

197



198

199 Asterisk indicates significant difference from Reference site (Kruskal-Wallis; Dunn; $p < 0.05$).

200

201

202

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205

206 Table 2 depicts the results of the multivariate test, which was performed to provide a
207 comprehensive understanding on the dependency relationships between exposure variables and
208 biological response markers. The canonical correlation analysis revealed the existence of a
209 significantly positive association ($R = 0.9968$; $p < 0.001$) between the set of environmental data
210 [X_i : Traffic (X1); and particulate matter fractions of PM_{10} (X2); $PM_{2.5}$ (X3); PM_{10} (X4)] and the
211 set of biological data [Y_i : leaf concentration of Cr (Y1) and Cd (Y2); micronucleus frequencies
212 (Y3) and pollen abortion (Y4)]. In a complementary way, a matrix containing the results of
213 Pearson's correlation tests between all variables used in the multivariate analysis is shown in
214 Table 3. Accordingly, there was significance between all combinations of correlations
215 considering environmental exposure variables and biological effects.

216 The present investigation was undertaken to continue a biomonitoring program of air
217 quality in areas of intense vehicular traffic, in the city of Uberlândia, Brazil, performed since
218 2007 (Pereira; Campos Júnior; Morelli, 2013). As an alternative to the assessment models
219 previously applied (Pereira et al., 2014; 2017), characterized by the *ex situ* exposure of plants
220 of the species *T. pallida* (previously cultivated in a reference site) to areas with high volume of
221 vehicle traffic, in this experiment, we investigated the relationship between environmental
222 variables (daily volume of vehicular traffic and concentration of particulate matter in the
223 atmosphere in different aerodynamic fractions) and biological biomarkers (metal
224 bioaccumulation, micronuclei formation and occurrence of pollen abortion) in response to plant
225 samples naturally exposed to air pollutants in permanent flowerbeds at the evaluated sites.

226 As evidenced by the results, the environmental concentrations of particulate matter
227 detected in this study do not exceed the average daily limits established by WHO guidelines
228 (2005) for preservation of the environmental quality and health of the exposed populations.
229 WHO parameters limited coarse particulate matter (PM₁₀) concentrations to 20 µg/m³ based on
230 an annual mean, and 50 µg/m³ based on a 24-hour mean; and also limited fine particulate matter
231 (PM_{2.5}) concentrations to 10 µg/m³ based on an annual mean, and 25 µg/m³ based on a 24-hour
232 mean.

233 Moreover, when there are differences in the concentration of Cd and Cr metals between
234 the leaf samples collected at the intersections and the reference site (approximately 4 x greater
235 in high-traffic locations), (i) the ability of *T. pallida* in bioaccumulating these contaminants and
236 (ii) the hypothesis that the particulate material from vehicle emissions carries metals are
237 confirmed. Furthermore, considering that metals adhered to particulate material may
238 contaminate plant species by direct entry via stomata or / and through the root system when
239 sedimented in the substrate (Campos et al., 2016; Kardel et al. 2018; Alatou, & Sahli 2019), the
240 observed genotoxic effects confirm that, even in concentrations below the standard limits
241 established by environmental legislation, the particulate material causes negative impacts on
242 biological systems, as verified by Domingues et al. (2018).

243 Interestingly, early detection of a hazardous exposure situation can significantly prevent
244 the occurrence of negative biological effects (Mukhopadhyay et al. 2020). In this sense,
245 biomarkers of response have been intensively applied in biomonitoring programs of
246 environmental quality, with the aim of providing subsidies for the implementation of preventive
247 measures and control of exposure to contaminants in the environment, including the review of
248 legal parameters (Rai 2016; Gillooly et al. 2019). The findings of this study highlight the

249 sensitivity of the genotoxicity biomarkers observed, which revealed a high frequency of MN
250 and pollen abortivity in *T. pallida* inflorescences exposed to contaminated air in areas of intense
251 vehicular traffic, establishing a qualitative and quantitative association between the exposure
252 variables and the detected biological effects.

253 Indeed, the combination of Trad-MCN test and pollen grain abortivity test, which therein
254 indicates genotoxic and mutagenic effects, respectively, emphasize the potential application of
255 *T. pallida* in environmental biomonitoring programs. Specifically, the associations between the
256 variables considered in this research also point to the perspective of *in situ* monitoring from the
257 planting of sensitive species in strategic locations for assessing environmental quality.

258

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CAPÍTULO III

ARTIGO CIENTÍFICO EXPERIMENTAL

Título:

Integrated biological assessment of air pollution in urban areas using passive biomonitoring

25

26 **Keywords:** Bioaccumulation; Heavy metals; Genotoxicity; Environmental Health;
27 Ecotoxicology.

28

29 **Introduction**

30

31 Biological indicators of air quality have been used as an important complement to
32 physicochemical analysis, required by regulatory and environmental protection agencies
33 (Campos et al. 2019, Ramić et al. 2019).

34 Although the main atmospheric contaminants, such as particulate matter and
35 polluting gases (nitrogen dioxide, sulfur dioxide, ozone, and carbon monoxid) are monitored,
36 there is scarce knowledge about the effects of the association of these contaminants in
37 biological systems (Kumar et al., 2023).

38 Biomonitoring offers the possibility of evaluating the responses of exposed
39 organisms, also allowing the prediction of genotoxic effects (Araújo et al. 2021). The use of
40 species that are sensitive to changes in concentrations of pollutants released into the
41 environment is a crucial strategy for establishing environmental health surveillance programs
42 based on scientific evidence (Fang et al. 2022).

43 Permanent monitoring systems are essential for establishing environmental quality
44 parameters based on biological response thresholds (DeBord et al. 2015, Cofin et al. 2022).
45 In a previous study, the analysis of *Tradescantia pallida* samples, present in public
46 flowerbeds located at different intersections in the city of Uberlândia, Brazil, provided
47 evidence of genotoxicity dependent on the intensity of vehicle traffic (Campos et al., 2020).

48 In this study, we focused on the use of the *Tradescantia pallida* species as a
49 biomonitor in situations where traffic varied in the same location, due to business closure
50 measures implemented as a strategy to face the COVID-19 pandemic.

51 More specifically, we tested the hypothesis that indicators of genotoxicity in *T.*
52 *pallida* respond significantly to changes in vehicular traffic in the same location.

53 To test the formulated hypothesis, we collected data on vehicular traffic, particulate
54 matter, heavy metals in the soil and plant, in addition to the genotoxicity analysis carried out
55 through the micronucleus test. Our results were evaluated based on multivariate analysis with
56 the objective of proposing a prediction model (multiple linear correlation) representing that
57 plants exposed to conditions of intense emission of pollutants from vehicles accumulate
58 heavy metals and present a significant increase in relation to the basal levels of genotoxicity
59 (expressed as frequency of micronucleated tetrads in pollen grains of young inflorescences).

60

61 **Material and Methods**

62

63 **Study design**

64 Four sampling campaigns of particulate matter, soil and plants were conducted over
65 a period in which local businesses were closed/reopened due to the Covid-19 pandemic
66 (Figure 1). The collection sites were the same points already monitored in a previous
67 feasibility study (Campos et al., 2020). The collection sites were chosen based on crossings
68 with different vehicular traffic intensity, but with an equivalent number of light/heavy
69 vehicles.

70

71

72 **Biological material**

73 Samples of leaves and young inflorescences of *T. pallida* were collected to assess
74 heavy metal concentration and genotoxicity, respectively. We used the mother plants, kept at
75 the Federal University of Uberlândia (area with low vehicular traffic intensity) as a reference
76 group for determining the baseline parameters of bioaccumulation and spontaneous
77 frequency of micronuclei.

78

79 **Heavy metals in soil and plant**

80 We collected and analyzed surface soil samples (volume = 100 cm³; depth = 5 cm)
81 and *T. pallida* leaves collected at different intersections (in quintuplicate) to determine the
82 concentration of heavy metals Lead (Pb), Chromium (Cr), Nickel (Ni), Cadmium (Cd),
83 Copper (Cu), and Zinc (Zn). We analyzed the biological samples according to the
84 methodology described in previous works (Campos et al. 2016; 2020), by using atomic
85 absorption spectrometry in a graphite oven (GFAA), according to the U.S. Environmental
86 Protection Agency methods (2007). We submitted the soil samples to the acid digestion
87 procedure (HCl 37% and HNO₃ 70%, 3:1 v/v) according to the ISO 11466:1995 protocol
88 (ISO, 1995).

89

90 **Street-level traffic data and particulate material sampling**

91 To monitor the effects of the variation in vehicle flow on the emission of particulate
92 matter, we used daily data on the circulating fleet provided by the Municipal Traffic
93 Department. The data was corrected so that the traffic volume was adjusted to the different
94 directions and lanes of the roads, also considering peak hours, weekends, and holidays. The
95 average daily levels of particulate material in the PM_{1, 2.5, 10} fractions at each evaluated site

96 and period were obtained using a portable particulate matter sampler (Dusttrak DRX Aerosol
97 Monitor - calibrated for detection of aerosol in sampling volume fixed at 3.0L/min).

98

99 ***Tradescantia* micronucleus assay (Trad-MN)**

100 We performed the Trad-MN assay to evaluate the biological response of genotoxicity
101 based on the micronucleus frequency. Following the protocol established by Ma et al. (1994),
102 we used young inflorescences collected from each monitored location in quintuplicate, which
103 were fixed in 1: 3 acetic acid to 70% ethanol solution for 24 h and preserved in 70% ethanol.
104 At least 10 slides (300 tetrads per slide) per collection site were analyzed using optical
105 microscope (400 x magnification).

106

107 **Statistical analysis**

108 Initially, we applied the Shapiro-Wilk test to assess the normality of the studied
109 variables. We compared heavy metal concentrations (both in plants and soil) between the
110 different sites and between collection periods, using two-way Analysis of Variance
111 (ANOVA). To compare the frequencies of micronuclei in inflorescences of *T. pallida* between
112 the different sites, we used the Kruskal-Wallis, followed by Dunn tests.

113 To evaluate the relationship between environmental variables (MP fractions, fleet and
114 metal concentration in soil and plant) as predictors of the observed genotoxic effect (MN
115 frequencies), we performed a multiple regression analysis and adjusted the mathematical
116 model to represent only the variables that contributed more significantly to the induction of
117 MN (>75%). P values <0.05 were considered statistically significant for all analyzes.

118

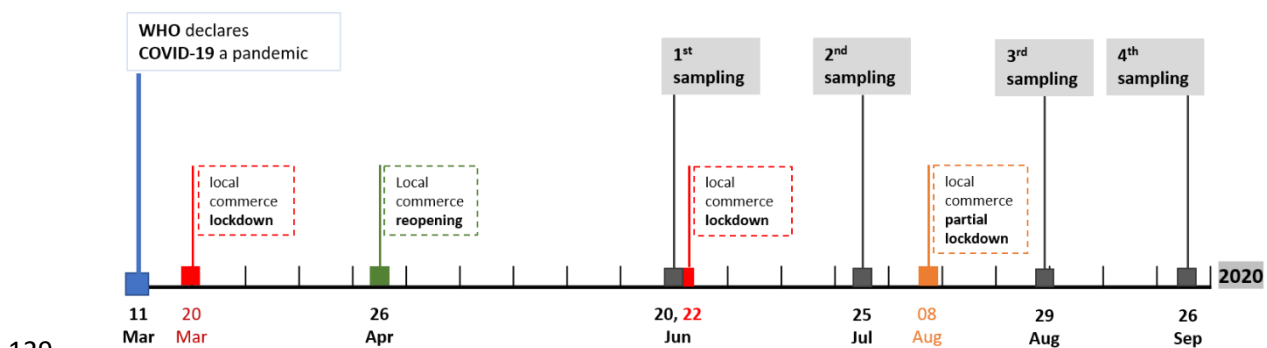
119

120 **Results**

121 According to the locations monitored during the four samplings carried out (Figure
122 1), it was possible to observe variations in the intensity of vehicular traffic, as can be seen in
123 Figure 2. The variations reflected measures to restrict commercial activities, imposed as a
124 result of the need control of the COVID-19 pandemic. In addition, it is possible to notice that
125 the concentration of particulate matter was also higher during the samplings carried out in
126 periods of greater intensity of vehicular traffic (first and fourth sampling).

127

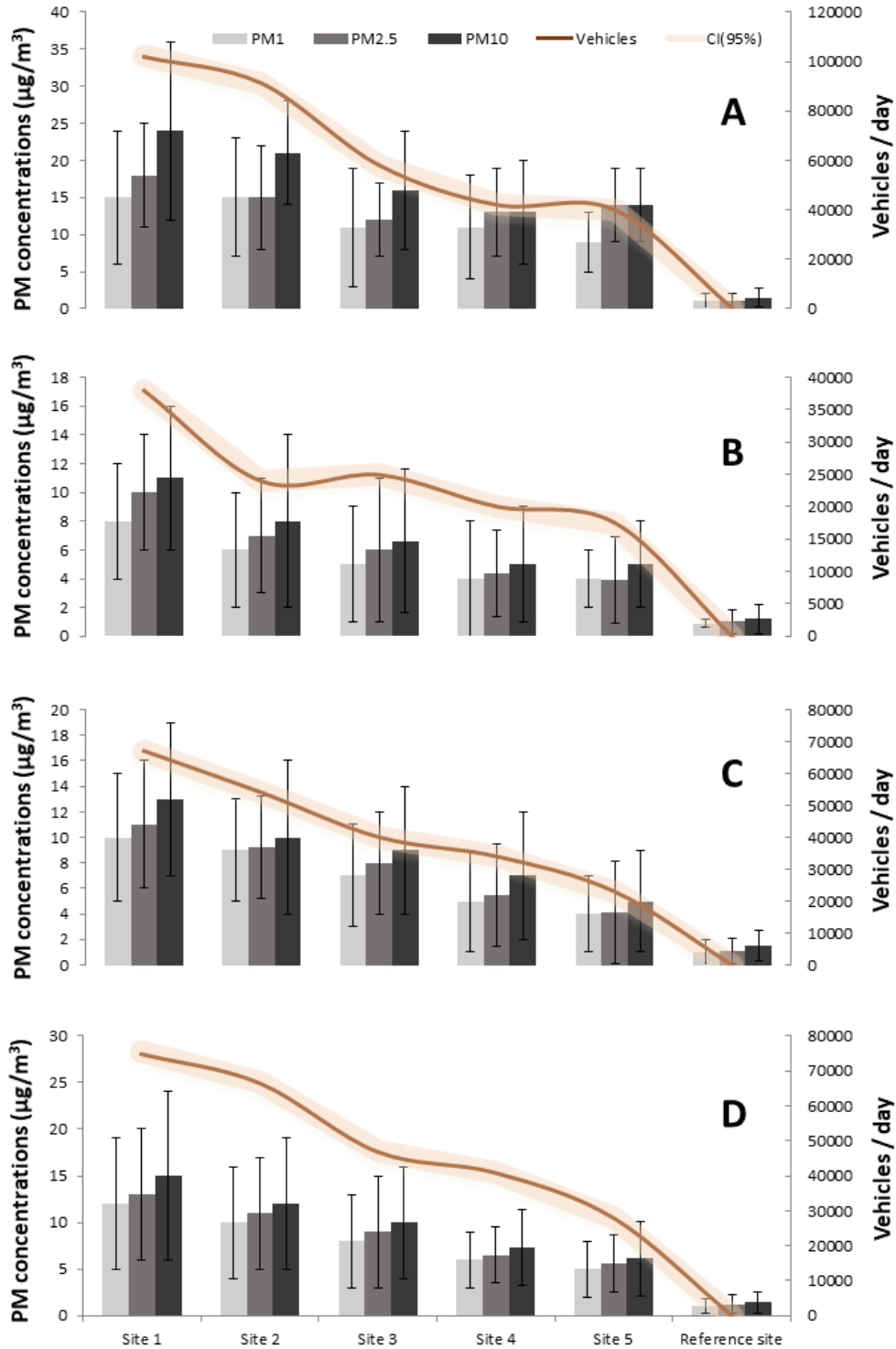
128 **Figure 1.** Sampling design of the study.



129

130

131 **Figure 2.** Fig. 2. Particulate Matter fractions (PM₁; 2.5; 10) concentrations and traffic volume
 132 (vehicles/day) in studied sites.



133

134 **Note:** Graphs A-D indicates 1st to 4th sampling according sampling design of the study.

135 **Table 1.** Metal analysis in soil and plants.

Site	[Metals (mg.kg ⁻¹)]												
	Sampling	Cd		Ni		Cr		Pb		Cu		Zn	
		Soil	Plant	Soil	Plant	Soil	Plant	Soil	Plant	Soil	Plant	Soil	Plant
1	1 st	0.4 ± 0.2	0.6 ± 0.2*	45.9 ± 31.1*	3.1 ± 2.4	60.2 ± 25.6*	2.3 ± 1.8*	8.5 ± 4.6	1.4 ± 0.9	29.3 ± 14.1	15.3 ± 7.8	96.0 ± 58.1	53.3 ± 15.4
	2 nd	0.1 ± 0.1	0.2 ± 0.1	24.7 ± 12.5	2.0 ± 1.0	10.7 ± 5.80	0.6 ± 0.4	5.7 ± 3.7	0.7 ± 0.6	28.2 ± 12.2	12.3 ± 5.4	77.2 ± 44.3	39.2 ± 24.9
	3 rd	0.2 ± 0.1	0.2 ± 0.1	18.5 ± 4.8	2.7 ± 1.8	19.5 ± 13.4	1.3 ± 0.9	5.6 ± 3.1	1.2 ± 0.7	18.3 ± 15.7	8.1 ± 5.2	82.1 ± 41.4	42.6 ± 31.3
	4 th	0.4 ± 0.2	0.5 ± 0.2	37.2 ± 23.6	4.0 ± 3.0	45.2 ± 25.8*	2.4 ± 2.0	10.0 ± 6.6	1.7 ± 1.1	23.5 ± 22.0	18.0 ± 10.1	76.4 ± 55.0	39.1 ± 24.0
2	1 st	0.4 ± 0.2	0.6 ± 0.3*	30.1 ± 11.3	2.2 ± 1.5	46.7 ± 16.3*	1.6 ± 0.8*	11.1 ± 5.2*	1.5 ± 1.0	19.5 ± 12.4	10.1 ± 4.9	85.5 ± 38.2	48.6 ± 24.7
	2 nd	0.1 ± 0.1	0.3 ± 0.2	14.2 ± 7.3	1.1 ± 0.5	14.4 ± 6.6	0.5 ± 0.3	8.9 ± 5.6	1.5 ± 1.1	16.7 ± 10.1	8.4 ± 4.3	87.2 ± 51.8	50.0 ± 23.9
	3 rd	0.1 ± 0.1	0.3 ± 0.2	18.1 ± 7.9	1.7 ± 1.2	13.3 ± 7.2	1.1 ± 0.9	7.5 ± 6.4	1.2 ± 0.9	15.7 ± 11.0	9.3 ± 4.7	72.7 ± 37.4	36.9 ± 19.4
	4 th	0.5 ± 0.3	0.7 ± 0.4*	30.0 ± 13.2	2.0 ± 1.7	35.1 ± 15.8*	1.2 ± 1.0	9.4 ± 7.0	1.6 ± 1.1	16.6 ± 12.2	11.6 ± 6.2	89.0 ± 50.2	49.4 ± 22.3
3	1 st	0.5 ± 0.4	0.7 ± 0.3*	28.2 ± 10.2	2.4 ± 1.4	35.6 ± 14.1*	1.1 ± 0.4	7.4 ± 3.6	1.1 ± 0.6	15.3 ± 7.2	8.2 ± 3.3	67.1 ± 26.7	37.3 ± 22.0
	2 nd	0.3 ± 0.1	0.4 ± 0.3	16.7 ± 9.4	1.5 ± 0.9	11.0 ± 3.4	0.7 ± 0.3	6.0 ± 4.1	1.0 ± 0.5	13.4 ± 5.5	7.6 ± 4.1	57.4 ± 20.8	30.2 ± 15.4
	3 rd	0.3 ± 0.2	0.4 ± 0.3	21.3 ± 11.0	1.9 ± 1.2	9.3 ± 4.5	1.0 ± 0.5	8.2 ± 5.1	1.2 ± 0.6	11.9 ± 6.0	6.6 ± 3.7	59.6 ± 19.2	28.8 ± 13.5
	4 th	0.6 ± 0.2*	0.8 ± 0.4*	33.2 ± 15.6	2.5 ± 1.6	33.3 ± 16.2*	1.2 ± 0.6	8.2 ± 4.5	1.2 ± 0.6	14.1 ± 6.3	7.6 ± 5.8	69.0 ± 30.0	39.5 ± 25.6
4	1 st	0.3 ± 0.1	0.3 ± 0.3	21.1 ± 7.6	2.1 ± 0.7	22.7 ± 10.2*	1.0 ± 0.4	7.5 ± 4.2	0.9 ± 0.5	5.9 ± 3.2	3.2 ± 1.4	60.2 ± 33.2	34.1 ± 17.5
	2 nd	0.2 ± 0.1	0.4 ± 0.3	12.3 ± 6.6	1.3 ± 0.8	8.3 ± 4.1	0.5 ± 0.2	7.1 ± 4.3	0.7 ± 0.4	7.2 ± 4.4	3.5 ± 1.3	48.0 ± 24.4	27.3 ± 13.8
	3 rd	0.2 ± 0.1	0.3 ± 0.3	10.5 ± 5.8	1.0 ± 0.5	7.6 ± 3.6	0.6 ± 0.3	6.7 ± 4.6	0.4 ± 0.4	5.6 ± 3.1	2.7 ± 0.9	49.7 ± 29.1	31.6 ± 15.7
	4 th	0.3 ± 0.2	0.4 ± 0.2	14.6 ± 7.7	1.4 ± 0.7	24.8 ± 9.0*	1.1 ± 0.6	8.4 ± 3.7	0.8 ± 0.4	6.4 ± 3.7	2.8 ± 0.8	59.3 ± 23.6	36.8 ± 22.0
5	1 st	0.4 ± 0.2	0.5 ± 0.3	23.7 ± 10.2	2.3 ± 0.9	14.5 ± 6.2*	0.9 ± 0.5	7.7 ± 3.7	1.1 ± 0.5	7.5 ± 4.4	4.0 ± 2.1	77.4 ± 28.6	46.4 ± 19.4
	2 nd	0.2 ± 0.1	0.2 ± 0.2	9.8 ± 5.5	0.9 ± 0.5	7.9 ± 5.0	0.4 ± 0.3	4.3 ± 2.8	0.6 ± 0.4	6.6 ± 3.9	4.5 ± 2.2	57.8 ± 32.1	29.9 ± 17.3
	3 rd	0.2 ± 0.1	0.3 ± 0.2	10.4 ± 6.1	1.1 ± 0.6	6.6 ± 4.7	0.7 ± 0.3	5.3 ± 3.1	0.6 ± 0.4	6.6 ± 3.1	4.2 ± 2.5	61.6 ± 40.7	38.6 ± 28.0
	4 th	0.3 ± 0.1	0.4 ± 0.2	17.1 ± 9.8	1.9 ± 0.8	15.3 ± 6.8*	1.0 ± 0.7	6.2 ± 4.0	0.9 ± 0.4	8.4 ± 3.7	5.6 ± 3.1	67.5 ± 31.4	44.2 ± 31.2
R	1 st	0.1 ± 0.1	0.1 ± 0.1	11.6 ± 5.6	1.4 ± 0.7	4.9 ± 2.3	0.3 ± 0.2	3.1 ± 1.1	0.4 ± 0.1	10.2 ± 5.2	5.3 ± 3.4	80.3 ± 34.4	52.3 ± 22.4
	2 nd	ND	0.1 ± 0.1	12.3 ± 6.1	0.8 ± 0.6	3.5 ± 3.1	0.3 ± 0.2	1.5 ± 0.9	ND	11.1 ± 6.1	5.6 ± 4.6	62.3 ± 26.6	34.5 ± 21.7
	3 rd	0.1 ± 0.1	0.1 ± 0.1	7.9 ± 4.4	0.9 ± 0.5	5.0 ± 2.0	0.4 ± 0.3	2.1 ± 1.2	0.3 ± 0.1	7.8 ± 5.4	5.0 ± 2.7	89.7 ± 33.3	44.7 ± 32.2
	4 th	ND	ND	8.8 ± 7.0	1.1 ± 0.8	5.6 ± 2.6	0.2 ± 0.2	2.0 ± 1.3	0.4 ± 0.1	9.6 ± 4.9	4.9 ± 2.9	67.2 ± 36.1	40.1 ± 31.0

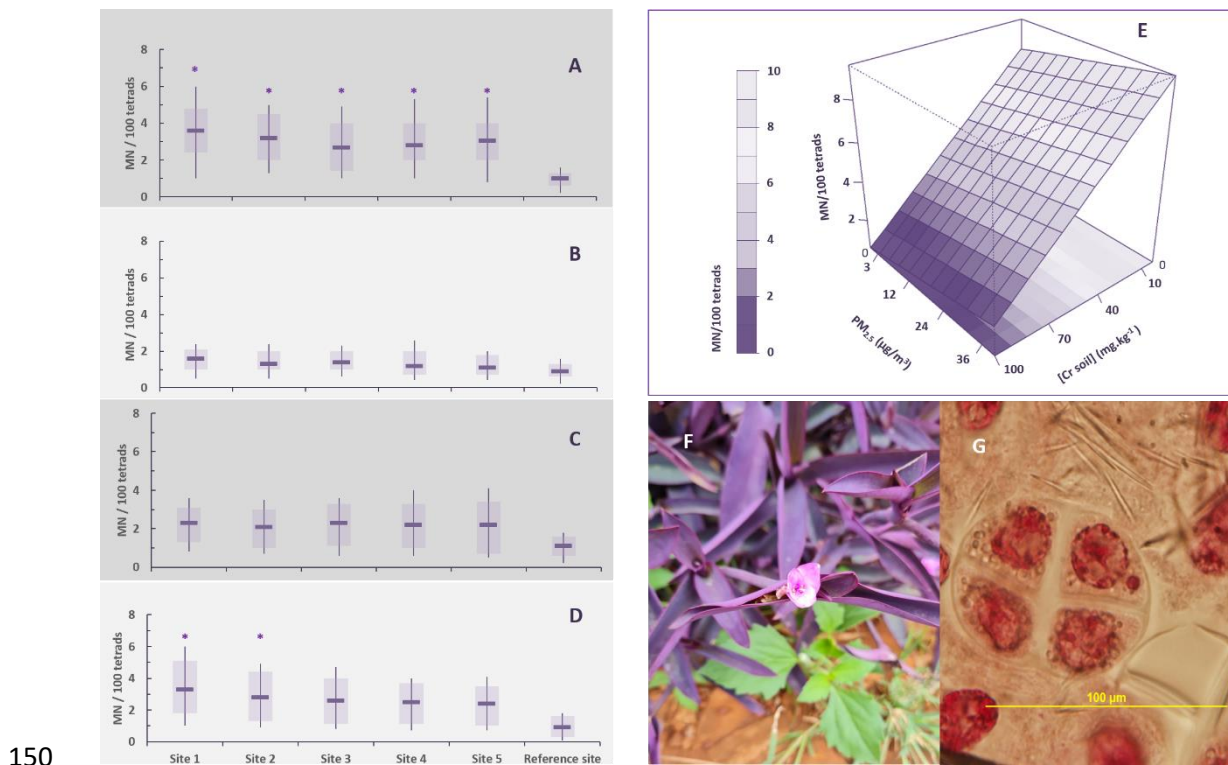
136 **Note:** Gray collar background indicates difference between sampling in a same site; *(Asterisk) indicates significant difference from reference
 137 site. Two-way Analysis of Variance (ANOVA).

138 According to the values depicted in Table 1, the concentrations of Cd and Cr were
139 significantly higher in soil and plant samples collected in sites of greater commercial activity,
140 during the period after the reopening of trade, indicating an effect of bioconcentration of metals
141 heavy in *T. pallida* - as a result of the transfer of contaminants between the atmosphere, soil and
142 biota compartments.

143 Figure 3 illustrates the results of the analyzes obtained from the Trad-MN assay. The
144 frequency of micronuclei observed in plants collected in commercial areas during the first and
145 fourth samplings was significantly higher compared to the values observed in the reference site,
146 which corresponds to the basal rate of micronuclei for the biomonitor plant.

147

148 **Figure 3.** Analysis of frequency of the micronuclei in *Tradescantia pallida* collected in
149 monitoring and reference sites.



150

151 **Note:** Micronucleus frequency (MN/100 tetrads) observed from locations 1st to 4th sampling (A-
152 **D)** *(Asterisk) indicates significant difference from reference site according Kruskal-Wallis
153 analysis, followed by Dunn tests. Graphical display of selected variables from a multiple
154 regression analysis (E). *Tradescantia pallida* var. *purpurea* (Rose) D.R.Hunt (F). Micronucleus
155 (MN) in early pollen tetrad cell (G).

156

157 From the mathematical model of multiple linear regression, we verified that genotoxicity
158 can be estimated with 79.72% of response dependent on the variables X1 (particulate matter in
159 the atmosphere – PM2.5) and X2 (concentration of chromium in the soil), which added,
160 respectively 77.50% and 2.22% to the coefficient of determination (R2). Therefore, MN
161 frequency (Y) can be predicted by the linear combination of these two variables (X1 and X2)
162 by the mathematical model: $[\log_{10} \text{MN} = 1.1033 + 0.0001X1 + (0.0222X2), R2 = 0.91,$
163 $F=17.07, p < 0.0001]$.

164

165 **Discussion**

166

167 Intense urbanization, accompanied by the maintenance of a vehicular fleet
168 predominantly powered by fossil fuels, results in deterioration of air quality, an important cause
169 of human illness in urban areas with intense vehicular traffic (Gao et al. 2023, Qi et al. 2023).

170 Vehicle emissions contain contaminants that, isolated or combined, produce genotoxic
171 effects on exposed organisms (Campos et al., 2020, Piccini et al., 2023, Qi et al 2023). Thus,
172 vehicle pollution is directly related to increased genomic instability, contributing to a higher
173 prevalence of chronic diseases.

174 The analysis of soil and air contamination is relevant and justified, especially as it allows
175 exposure via the food chain (Kumar et al. 2019), since microorganisms and plants accumulate
176 pollutants such as heavy metals (Wai et al. 2017, Sharafi et al. 2022).

177 Furthermore, the integrated monitoring of pollutants in the air, soil, and sentinel plants
178 allows to assess the persistence of genotoxic agents in the environment (Érseková et al. 2014).
179 Our findings showed how the changes in vehicle traffic promote effects on the atmospheric
180 concentration of particulate matter. We also demonstrate that the suspension of pollutant
181 emission sources alters the concentrations of metals in the environment but does not completely
182 remove them, since chemicals, especially heavy metals, are transferred between different
183 environmental compartments, such as by the air-soil-plant (Chen et al. 2019, Moradi et al. 2021).
184 Our results are compatible with the few studies that evaluated the genotoxicity induced by
185 atmospheric environmental contamination from vehicles in accumulator plants (Badran et al.
186 2020). Previous reports have shown the accumulating potential of *T. pallida* for Cr (Sinha et al.
187 2017), Cd and Pb (Campos et al., 2016), and Mn (Amato-Lourenço et al. 2017).

188 We also observed that the Cr bioconcentration was significantly higher in collection sites
189 and periods with higher traffic intensity and particulate matter emission - the participation of Cr
190 in the oxidative processes related to genotoxicity effects has already been reported in the
191 literature (Kapoor et al. 2022). The genotoxic effect can be monitored from the visualization of
192 micronuclei, which result from chromosomal losses or breaks, indicating, respectively,
193 aneugenic and clastogenic responses (Sharma et al. 2012).

194 Our results demonstrate the concentration-dependent relationship between particulate
195 matter and micronucleus frequency in *Tradescantia pallida*. Here, it is important to highlight
196 that the maintenance of *T. pallida* matrices to recompose the beds (monitoring sites) was an

197 important factor to guarantee isogenicity in evaluating genotoxic effects, so that the responses
198 can be attributed to environmental variations.

199 Thus, our results not only validate the sensitivity of *T. pallida*, already evaluated in other
200 studies (Santos et al., 2015, Campos et al. 2020, Khosrovyan et al., 2022), but also confirm the
201 hypothesis that the plant responds significantly to changes in vehicular traffic in the same
202 environment, based on indicators of genotoxicity.

203 Our data reinforce the recent understanding that air quality assessments are incomplete
204 when the focus is on the physicochemical analysis of contaminants. It is important to evaluate
205 the effects of complex mixtures that occur in the environment and cause impacts on exposed
206 organisms, such as the genotoxic effects evidenced in our study.

207 Based on the mathematical model presented, our findings report that continuous
208 monitoring of environmental exposure indicators (atmospheric and soil contamination) and
209 effects (bioconcentration of metals and genotoxicity) can be used as a surveillance and warning
210 system about the risks of atmospheric pollution. Additionally, the results obtained are scientific
211 evidence, supporting decision-makers about investments, control, inspection, and other
212 measures that improve the quality of life of urban populations.

213

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